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## **Reducing Petroleum Consumption from Transportation**

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# Reducing Petroleum Consumption from Transportation

Christopher R. Knittel\*

## Abstract

The United States consumed more petroleum-based liquid fuel per capita than any other OECD-high-income country—30 percent more than the second-highest country (Canada) and 40 percent more than the third-highest (Luxemburg). This paper examines the main channels through which reductions in U.S. oil consumption might take place: (a) increased fuel economy of existing vehicles, (b) increased use of non-petroleum-based low-carbon fuels, (c) alternatives to the internal combustion engine, and (d) reduced vehicles miles travelled. I then discuss how the policies for reducing petroleum consumption used in the US compare with the standard economics prescription for using a Pigouvian tax to deal with externalities. Taking into account that energy taxes are a political hot button in the United States, and also considering some evidence that consumers may not correctly value fuel economy, I offer some thoughts about the margins on which policy aimed at reducing petroleum consumption might usefully proceed.

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# 1 Introduction

The United States consumed more petroleum-based liquid fuel per capita than any other OECD-high-income country—30 percent more than the second-highest country (Canada) and 40 percent more than the third-highest (Luxemburg). The transportation sector accounts for 70 percent of U.S. oil consumption and 30 percent of U.S. greenhouse gas emissions. Gasoline and diesel fuels, alone, account for 60 percent of oil consumption. The economic argument for seeking to reduce this level of consumption of petroleum-based liquid fuel begins with the externalities associated with high levels of U.S. consumption of petroleum-based fuels.

First, burning petroleum contributes to local pollution. The transportation sector accounts for 67 percent of carbon monoxide emissions, 45 percent of nitrogen oxide (NO<sub>x</sub>) emissions, and 8 percent of particulate matter emissions. These pollutants lead to health problems ranging from respiratory problems to cardiac arrest. Furthermore, automobiles emit both NO<sub>x</sub> and volatile organic compounds which, combined with heat and sunlight, form ground-level ozone, or smog. [Currie and Walker \(2011\)](#) and [Knittel, Miller, and Sanders \(2011\)](#) both find that decreases in traffic reduce infant mortality.

Second, burning a gallon of gasoline causes roughly 25 pounds of carbon dioxide to be emitted into the atmosphere, which raises the risks of destructive climate change. [Greenstone, Kopits, and Wolverton \(2011\)](#) estimate the social cost of carbon under a variety of assumptions. They estimate an SSC as high as \$65 per metric ton of carbon dioxide (and gases with an effect equivalent to carbon dioxide) in 2010, but values that are often in the range of \$21 to \$35 per metric ton. [Tol \(2008\)](#) is a meta-study of the existing literature, for studies written after 2001, he finds the median social cost of carbon ranges from \$17 to \$62 per metric ton of CO<sub>2</sub>-equivalent.

The U.S. dependence on imported gasoline has had other costs, too, including the military expense of trying to assure stability in oil-producing regions (for example, [ICTA \(2005\)](#)), and the relationship between oil price shocks and macroeconomic downturns. These, too, can be viewed as negative externalities. But this paper neither focuses on these various externalities and social costs, nor delves into the literature about quantifying them. Instead, I take their existence as largely given, and instead focus on understanding the policy tools that seek to reduce gasoline consumption.

Of course, an obvious starting point for economists is to look at prices: although the price of petroleum is set in a global market, government taxes on petroleum vary quite substantially. Table 1 lists taxes on gasoline and diesel on a per gallon basis as of 2010 for OECD Category I countries—essentially the worlds most developed economies. The United States and Canada are clearly outliers,

with taxes on gasoline below \$1 per gallon. How do these price differences affect consumption? Figure 1 is suggestive. It plots the per capita petroleum-based liquid fuel consumption for these countries versus the gasoline price in that country, with the size of the bubbles are proportional to population. The regression line is population-weighted, but looks similar if it is not weighted. It would require quite a bit of additional argument and delicacy to estimate a reliable elasticity of demand from this data, but for the record, the slope of fitted log-log regression line through these data is -1.86. If one were to also include the log of income as an explanatory variable in such a regression, the coefficient associated with the log of gasoline prices is -1.49, while the coefficient associated with the log of income is 1.05.

The relative fuel use across the United States and other OECD Category I countries is, at least in part, a by-product of differences in the types and use of light-duty vehicles. Schipper (2006) reports that the average gallons-per-mile of European fleets in 2005 was below 0.034 (29.4 miles-per-gallon), while the average gallons-per-mile of the U.S. fleet was 0.051 (19.6 miles-per-gallon). Because of differences in how fuel efficiency is evaluated, this finding probably understates the European advantage. Similarly, Schipper reports that per capita miles travelled in European countries is between 35 to 45 percent of U.S. miles travelled.

The next four sections of this paper examine the main channels through which reductions in U.S. oil consumption might take place: (a) increased fuel economy of existing vehicles, (b) increased use of non-petroleum-based low-carbon fuels, (c) alternatives to the internal combustion engine, and (d) reduced vehicles miles travelled. I then discuss how these policies for reducing petroleum consumption compare with the standard economics prescription for using a Pigouvian tax to deal with externalities. Taking into account that energy taxes are a political hot button in the United States, and also considering some evidence that consumers may not correctly value fuel economy, I offer some thoughts about the margins on which policy aimed at reducing petroleum consumption might usefully proceed.

## 2 Improved Fuel Economy

Shortly after the oil prices shocks of the 1970s, the United States adopted Corporate Average Fuel Economy (CAFE) standards, which set minimum average fuel economy thresholds for the new vehicles sold by an automaker in a given year. Figure 2 shows how the standard evolved. For passenger cars, the standard increased by only 0.5 miles-per-gallon from 1984 to 2010; for light-duty trucks, the increase was only 3.5 miles-per-gallon over this same time period. From 1978 to 1991 the

standard for light trucks differentiated between two- and four-wheel drive trucks, but manufacturers could also choose to meet a combined-truck standard. By world standards, these miles-per-gallon standards are not aggressive. After accounting for differences in the testing procedures, the World Bank estimated that the European Union standard was roughly 17 miles-per-gallon more stringent in 2010 than the U.S. standard (An, Earley, and Green-Weiskel (2011)).

Manufacturers who violate the CAFE standard pay a fine of roughly \$50 per mile-per-gallon per vehicle for violating the standard. Historically, U.S. manufacturers have complied with the standard. Asian manufacturers have typically exceeded the standard in each year, while European manufacturers have typically violated the standard and paid the fines. Trading between manufacturers was not allowed, so there was no possibility for certain manufacturers to accumulate credits for selling a higher proportion of fuel-efficient cars and then selling those credits to other manufacturers.

Other than the fact that the standards have barely budged over the last three decades, two features of the original CAFE standards reduced their effect. First, sport-utility vehicles were treated as light trucks, and thus could meet a lower miles-per-gallon standard than cars. Perhaps not coincidentally, in 1979 light trucks comprised less than 10 percent of the new vehicle fleet, but this share rose steadily and peaked in 2004 at 60 percent. Second, vehicles with a gross vehicle weight of over 8,500 pounds, which includes many large pickup trucks and sports-utility vehicles, were exempt from CAFE standards.

Actual new vehicle fleet fuel economy in the United States has changed little since the early 1980s. Figure 3 plots the fuel economy of passenger vehicles (cars) and light duty trucks from 1979 to 2011. The figure shows that while the average fuel economy of both cars and trucks increased over this time period, fleet fuel economy fell as consumers shifted away from cars and into trucks. The figure also shows that during the run-up in gasoline prices beginning in 2005, fleet fuel economy increased. This rise appears to have subsided by 2010.

Although the fuel economy of new U.S. vehicles gradually declined through the 1980s and 1990s, there was scope for substantial improvements. In the short run, when the set of offered vehicles is fixed, car-buyers could choose vehicles with higher fuel efficiency. In 2011, for example, while the mean passenger car available for sale was rated at 23 miles-per-gallon, 10 percent of passenger cars had a rating of 30 miles-per-gallon or above. The highest rating for 2011 was the Nissan Leaf at 99 miles-per-gallon; the Toyota Prius had a combined fuel economy rating of 50 miles-per-gallon.

In the medium run, automakers can adjust vehicle attributes by trading off weight and horsepower for increased fuel economy. Knittel (2011) finds that reducing weight by 1 percent increases

fuel by roughly 0.4 percent, while reducing horsepower and torque by 1 percent increases fuel economy by roughly 0.3 percent

In the long run, manufacturers can push out the frontier. [Knittel \(2011\)](#) estimates that had manufactures put all of the technological progress observed in the market from 1980 to 2006 into fuel economy, instead of putting it into attributes that increased horsepower and/or weight, average fuel economy would have increased by 60 percent, instead of the 11.6 percent increase actually observed. On average, a vehicle with a given weight and engine power level has a fuel economy that is 1.75 percent higher than a vehicle with the same weight and horsepower level from in the previous year. While the analysis in [Knittel \(2011\)](#) ends in 2006, using similar data and empirical models through model year 2011, the technological frontier has shifted out at an average rate of 1.97 and 1.51 percent per year from 2006 to 2011 for passenger cars and light-duty trucks, respectively, suggesting that no technological barrier has yet been reached. The greater availability of hybrids and plug-in hybrids also suggest that progress is likely to continue.

Gasoline prices do seem to affect choices about which cars to buy. A number of papers have used this variation in gasoline prices to estimate the magnitude of this response. These papers inevitably estimate a short run response to gasoline prices—in the particular sense that the choice set of vehicles is usually held fixed. [Busse, Knittel, and Zettelmeyer \(2009\)](#) estimate that over the period of 1999 to 2008, the market share of vehicles in the bottom quartile of fuel efficiency, among those vehicles offered in a given year, falls by nearly 24 percent for every \$1 increase in gasoline prices. In contrast, the market share of the upper quartile of vehicles as ranked by fuel efficiency increases by over 20 percent. They also show that the market share of compact cars increases by 24 percent for every \$1 increase in gasoline prices, while the market share of sport-utility vehicles falls by 14 percent. [Klier and Linn \(forthcoming\)](#) estimate a logit demand system and focus on the effects of changes in a vehicles cost-per-mile on demand. They find that a 5-cent increase in a vehicles cost per mile, equivalent to a one-dollar increase in gasoline prices for a 20 miles-per-gallon vehicle, decreases the log of its market share by between 0.5 and 0.8, all else equal.<sup>1</sup> In the aggregate, this translates to an increase in average fuel economy of between 0.5 and 1.2 miles-per-gallon for every \$1 dollar increase in gas prices. Again, their estimates hold the set of offered vehicles fixed. [Li, Timmins, and von Haefen \(2009\)](#) find similar effects.

A new Corporate Average Fuel Economy standard in place for 2011 seeks to increase average fuel economy to roughly 34.1 miles-per-gallon by 2016. The Environmental Protection Agency

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<sup>1</sup>These coefficients represent the change in a vehicles log of market share when the vehicles cost per mile increases. Of course, if this is driven from changes in gasoline prices, then the cost of all vehicles cost per mile will change. This explains why these effects are so large.

and Department of Transportation are currently in the rulemaking process for model years 2017 and beyond, with President Obama and 13 automakers agreeing to a standard of 54.5 miles-per-gallon by 2025. A number of notable changes have occurred. First, the mileage standards are now based to some extent on the greenhouse gas emissions of the vehicle, which can deviate from fuel economy because of ancillary greenhouse gas emissions associated with, for example, air conditioner refrigerant leaks. Second, the new standards are footprint-based, in which each vehicle face a standard based on the area of the footprint of its tires; larger footprints face a lower standard. For example, the 2011 Honda Civic coupe has a footprint of 43 square feet, while the 2011 Ford F-150 SuperCab has a footprint of 67 square feet. In 2016, these vehicles would face fuel efficiency standards of 41.1 miles-per-gallon and 24.7 miles-per-gallon, respectively.<sup>2</sup> For more details on the fuel efficiency rules for the next few years, see [U.S. Energy Information Administration \(2007\)](#) and [U.S. Environmental Protection Agency \(2010\)](#).

Is new vehicle fuel economy of 34.1 and 54.5 miles-per-gallon in 2016 and 2025, respectively, attainable? If we take the average rates of technological progress from [Knittel \(2011\)](#) and a new vehicle fuel economy in 2010 of roughly 29 miles-per-gallon, new vehicle fuel economy in 2016 would be roughly 32 miles-per-gallon in 2016, close to the standard of 34.1 miles-per-gallon. Using the estimated trade-off coefficients, getting to 34.1 miles-per-gallon would require reducing weight and engine power by less than 6 percent. Alternatively, increasing the rate of technological progress to 2.75 percent per year would achieve the mark.

And, what about the 2025 standard of 54.5 miles-per-gallon in 2025. Taken literally, it would require fundamental changes to rates of technological progress and/or the size and power of vehicles. The 2025 number is a bit misleading. In the law, the 54.5 miles-per-gallon standard is based on a calculation from the Environmental Protection Agency based on carbon dioxide tailpipe emissions. It also includes credits for many technologies, including plug-in hybrids, electric and hydrogen vehicles, improved air conditioning efficiency; and others. On an apples-to-apples basis, [Roland \(2011\)](#) cites some industry followers that claim that the actual new fleet fuel economy in 2025 is more like to 40 miles-per-gallon. Achieving 40 miles-per-gallon by 2025 is certainly possible. At a rate of technological progress of 1.75 percent per year, 40 miles-per-gallon requires additional reductions in weight and engine power of less than 7 percent.

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<sup>2</sup>See, <http://www.nhtsa.gov/cars/rules/cale/overview.htm>. The sticker fuel economy is roughly 80 percent of how the vehicle is counted for the CAFE standard.

### 3 Alternative Fuels

Biofuels are derived from biological components like corn, soybeans, sugar, grasses, and wood chips. The ethanol produced by this process is an imperfect substitute for gasoline, although bio-diesel as a nearly perfect substitute for petroleum-based diesel. (Methanol is another alcohol and imperfect substitute for gasoline that can be derived from either methane—that is, natural gas—or coal.) Biofuels also hold the potential to have lower carbon emissions. If the plant material could be grown and converted to liquid fuel using only technologies that do not produce any greenhouse gas emissions and not lead to land use changes that increase greenhouse gases, then biofuels would not emit any net greenhouse gases over the life-cycle.

In practice, the lifecycle emissions of biofuels are affected by a number of factors. First, the feedstock used affects carbon emissions during the growing stage—for example, through fertilization. The most common feedstock used in the U.S. is corn. Brazilian ethanol is made from sugar cane. So-called second generation or cellulosic ethanol uses feedstocks that require little in the way of irrigation and fertilizer during the growing process, such as miscanthus and switchgrass. Second, the fuel used for generation of heat and electricity during refining process affects emissions. Third, the calculation is affected by whether the co-products from distilling, notably distillery grains with solubles, are dried before being sold and whether these emissions should be included, or treated as another product.

Fourth, the life-cycle emissions of corn-based biofuels are affected by the milling process. Corn ethanol is typically refined using either a dry or wet milling process. Under wet milling, the corn is soaked in hot water and sulfurous acid. The starches from this mixture are then separated and fermented, leading to ethanol. Dry milling requires less energy and generates fewer greenhouse gas emissions, but does not yield as many co-products as wet milling. Under dry milling, the corn is ground into flour and cooked along with enzymes, where yeast is added for fermentation. The ethanol is then separated from the liquid. The remaining component undergoes another process turning it into livestock feed.

Finally, and most difficult to estimate, increases in biofuel production can alter land use patterns elsewhere. For example, an increase in Brazilian sugar-cane ethanol may reduce pasturelands and thus cause cattle farmers to cut down rainforest, which reduces the quantity of greenhouse gases sequestered by the rainforest. The influential paper by [Searchinger et al. \(2008\)](#) was the first to measure this factor, finding that once indirect land use effects are considered, corn-based ethanol can have nearly twice the greenhouse gas emissions of gasoline. A number of follow up papers have found that while these effects may not be this large, they remain important. For example, [Tyner](#)



et al. (2010) argue that once changes in both international trade and crop yields are accounted for, corn ethanol results in fewer GHG emissions than gasoline, despite indirect land use changes.

How does the sum of these factors compare to the emissions of gasoline? The emissions of a gallon of gasoline over the entire life-cycle of its production depend on, amongst other things, the efficiency of the refinery and weight of the oil. A number of estimates exist. The [California Air Resource Board \(2011\)](#) estimated that an average gallon of California-refined gasoline generates 27.9 pounds of carbon dioxide-equivalent greenhouse gas emissions. Roughly 19 pounds of this comes from the combustion of the gasoline, while the remainder comes from the emissions associated with refining, transporting, and so on. The 19 pounds figure may sound too high, given that a gallon of gasoline weighs roughly 6 pounds. The reason is that during the combustion process the carbon atoms in the gasoline, which have a molecular weight of 12, combine with 2 oxygen atoms, each having a molecular weight of 16, from the atmosphere.

The [California Air Resource Board \(2011\)](#) also estimates that lifecycle emissions for a number of ethanol pathways lead to higher greenhouse gas emissions than gasoline. For example, Midwest ethanol (shipped to California) produced using a wet mill process and coal for heating and electricity has 26 percent more greenhouse gas emissions than the average gasoline refined in California. In contrast, dry mill, wet distillery grains with solubles Californian ethanol that uses 80 percent natural gas and 20 percent biomass is predicted to have greenhouse emissions that are 19 percent below that of gasoline. Brazilian ethanol made from sugarcane has the lowest life-cycle emissions among those pathways analyzed in the California report. An [U.S. Environmental Protection Agency \(2007\)](#) report reaches similar conclusions. Dry mill ethanol made using coal has either 13 or 34 percent more emissions than gasoline. However, dry mill ethanol using biomass, a form of cellulosic ethanol, in a combined heat and power system has 26 or 47 percent fewer emissions.

In short, lifecycle analyses suggest that corn-based ethanol can play only a marginal role in reducing greenhouse gas emissions from the transportation sector. In contrast, cellulosic-based biofuels can potentially play a much larger role, although there remain technological obstacles to widespread mass production of ethanol at low cost from this source.

There are other natural limits to the impact of corn-based ethanol production in the United States as well. How much farmland would be required if Americas cars were to run solely on E85: 85 percent ethanol and 15 percent gasoline? Well, gasoline usage in the United States is roughly 140 billion gallons per year, and it takes 128,500 acres of corn to produce 50 million gallons of ethanol (according to the FAQ at <http://ethanol.org>). Given that ethanol has an energy content that is roughly 67 percent of gasoline, 140 billion gallons of our current fuel, which is roughly 5

percent ethanol, would require roughly 190 billion gallons of E85. Thus, if the ethanol used corn as the feedstock, this would imply roughly 415 million acres of corn crop but there is currently only 406 million acres of farmed land in the United States. In short, significant expansion of corn-based ethanol production is likely to require additional land, which unleashes environmental consequences discussed earlier. In addition, corn-based biofuels also compete with current uses of corn, which has implications for the worldwide price of corn and other substitute grains. Cellulosic biofuels, in contrast, offer a feedstock that will not compete with food products nearly as much, since these plants can be grown on marginal lands and without irrigation.

Large scale substitution of ethanol for gasoline is limited in the short run because of the the blend wall—the percentage of fuel that can be ethanol and safely burned in a vehicle designed to burn only gasoline. The Environmental Protection Agency recently ruled that vehicles of model year 2005, or newer, can safely burn fuel that is 15 percent ethanol. Vehicles older than this can burn E10. Flex-fuel vehicles, in contrast, can burn fuel that is up to 85 percent ethanol.

U.S. policymakers have adopted a variety of bio-fuel policies: performance standards, subsidies, and mandates. The existing subsidy for ethanol is the Volumetric Ethanol Excise Tax Credit. The current credit offers fuel blenders \$0.45 tax credit per gallon of ethanol sold. Before this tax credit, ethanol received an implicit subsidy (relative to gasoline) as it was exempted from the federal fuel excise tax in 1978. The 2008 Farm Bill differentiated between corn-based and cellulosic ethanol, with cellulosic ethanol receiving a \$0.91-per-gallon tax credit, minus an applicable tax credit collected by the blender of the cellulosic ethanol.<sup>3</sup> Small ethanol producers—those with a capacity of less than 60 million gallons—receive an additional 10 cents per gallon credit.

Similar subsidies exist for biodiesel. The Jobs Creation Act of 2004 established a \$1-per-gallon tax credit for biodiesel created from virgin oil, defined as oil coming from animal fats or oilseed rather than recycled from cooking oil. Biodiesel from recycled oil receives a \$0.50-per-gallon tax credit. These subsidies were extended under the Energy Policy Act of 2007. The tax credits for both ethanol and diesel have been extended to the end of 2011.

The other major federal ethanol policy is mandates to use such fuels. The first Renewable Fuel Standard was adopted in 2005. The Energy Independence and Security Act of 2007 expanded this standard by calling for 36 billion gallons of bio-fuels—including 21 billion gallons of advanced

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<sup>3</sup>These figures understate the subsidy level because they are on a per-gallon basis, not on a per-energy basis. As noted in the text, one gallon of ethanol has roughly 67 percent of the energy content of a gallon of gasoline, implying that it requires 1.48 gallons of ethanol to displace one gallon of gasoline. Therefore, on a per gallon of gasoline equivalent basis, corn-based ethanol receives a 67 cents per gallon of gasoline equivalent subsidy; 81 cents for a small producer. Cellulosic ethanol receives a  $1.35 \text{ per gallon of gasoline equivalent subsidy}$ , 1.49 per gallon of gasoline equivalent for small producers.

bio-fuels by 2022, which are to have a lower greenhouse-gas content than corn-based ethanol. Given how the Renewable Fuel Standard is implemented, ethanol prices reflect an implicit subsidy, while gasoline is priced as if it were taxed (Holland et al. (2011)). A variety of state-level blend minimums and performance standards also exist.

Methanol is another alcohol that can be used as a liquid fuel. Methanol production is an established industry: methanol is used as a racing fuel, as an industrial chemical, and as a liquid fuel in some countries—especially China. Methanol can be produced from natural gas, coal, or wood. In 2010, the U.S. consumed 1.8 billion gallons of ethanol with world production totaling over 15 billion gallons (see statistics at <http://methanol.org>), roughly on par with global ethanol production of 23 billion gallons in 2010 (see statistics at <http://ethanolproducer.com>). In contrast to ethanol, most methanol consumption is not as a fuel, but as a chemical feedstock.

Methanol has three main advantages over corn-based ethanol. First, on a greenhouse gas basis, Delucchi (2005) estimates that methanol produced from natural gas has 11 percent lower greenhouse gas emissions than corn-based ethanol. However, he finds that methanol still has higher emissions than gasoline. Others find that the greenhouse gas emissions from methanol are roughly equivalent to gasoline (MIT Future of Natural Gas Study (2011)). Second, methanol is cheaper than gasoline, at least at current oil and gas prices. Methanex, the worlds largest methanol producer, quotes current retail methanol prices in North America of \$1.38 per gallon. Methanol has an even lower energy content than ethanol at roughly 53 percent of gasoline, implying a cost per gallon of gasoline equivalent of (approximately, because of changes in engine efficiency) \$2.51 per gallon. Third, methanol doesnt rely on crops, eliminating the negative consequences associated with crop production.

Methanol also faces four disadvantages. First, methanol produced from natural gas cannot achieve the same reductions in greenhouse gas emissions as second-generation or cellulosic ethanol. Second, alcohols are generally more corrosive than gasoline, and methanol is even more corrosive than ethanol. For vehicles to run on ethanol or methanol, manufacturers must protect certain engine parts and rubber material from the fuels. Flex-fuel vehicles that can run on fuel that is as much as 85 percent methanol (M85) require a slightly larger investment, on the order of \$200 per vehicle (MIT Future of Natural Gas Study (2011)). Third, as discussed above, methanol has an even lower energy content than ethanol, so a tank of gas wouldnt take you as far.<sup>4</sup> Finally, there are open questions as to how safe the drilling process is, or can be, for shale gas, including potential

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<sup>4</sup>Because of their lower vapor pressure, starting an engine in cold weather is more difficult when using ethanol and methanol (with ethanol having a lower vapor pressure compared to methanol), which may prompt consumers to use a lower blend of these fuels during the winter.

problems of methane leakage.

Given the recent discoveries of large shale gas deposits within North America, a compelling argument can be made that methanol, as a substitute for gasoline, should have the same support as corn-based ethanol. Methanol carries similar greenhouse gas reductions, if not larger, and is not petroleum based. The open issue is whether drilling for shale gas has fewer environmental repercussions than the land use implications of ethanol.

An alternative use for natural gas is in compressed natural gas (CNG) vehicles, which use internal combustion engines to burn natural gas stored at high pressures. [Werpy et al. \(2010\)](#) summarize tailpipe emission comparisons of vehicles and find that compressed natural gas has emission reductions that are often above 20 percent, compared to gasoline, but often less than 10 percent when compared to diesel fuel. Moreover, long run average costs on a gallon-of-gasoline equivalent are currently below gasoline: the U.S. Department of Energy reports national average prices of \$2.09 for October 2011.

The drawbacks to CNG vehicles are similar to electric vehicles (discussed below). New infrastructure is needed for refueling with compressed natural gas. Refueling can take longer, especially if done at home: slow-fill home units can take over four hours. CNG vehicles have limited range, often the equivalent of about eight gallons of gasoline. CNG vehicles also have a higher upfront cost: the Honda Civic GX, a CNG vehicle, sells for roughly a \$4,000 premium, but has 27 percent less horsepower than a comparable gasoline-powered car. A thorough comparison of CNG and electric vehicles is beyond the scope of this paper, but again, given large natural gas reserves recently discovered, this would appear to be a worthwhile avenue for research. My read of the literature suggests that these drawbacks are not as severe with CNG vehicles as with electric vehicles, although the reduction in carbon emissions from CNG vehicles may also be less than if the electricity for a vehicle is generated in a low-carbon manner. Once the benefits of both greenhouse gas emissions and petroleum reductions are compared with the added costs, CNG vehicles might make more sense than electric vehicles.

## 4 Replacing the Internal Combustion Engine

Shifting away from the internal combustion engine to powering vehicles with electricity or with hydrogen is another way of reducing petroleum usage. Either approach could represent a reduction in the pollutants per unit of energy of the fuel and an increase in fuel economy—as measured by the energy required to travel one mile. In terms of greenhouse gas emissions, all-electric or

hydrogen vehicles have been viewed by some as the end game, since it is possible to generate either electricity or hydrogen in a carbon-free way—say, through solar or wind power. Ultimately, both technologies would probably use electric motors. It is possible to burn hydrogen directly in an internal combustion engine. BMW, for example, has a flex-fuel 7-series that can use both diesel and hydrogen. However, this forgoes the efficiency gain from electric motors, so most industry followers believe that if hydrogen were to penetrate the market it would do so through a fuel cell that powered an electric-motor vehicle.<sup>5</sup>

The hurdle for both electricity and hydrogen technologies is, of course, cost. These costs can usefully be divided up into costs for vehicles, cost of the electricity or hydrogen itself, and infrastructure costs associated with re-energizing vehicles.

For a pure electric vehicle, battery technology still imposes some daunting constraints. While I am not aware of any studies detailing the required battery size as a function of key variables such as the vehicle weight and desired range, some rough calculations are possible. My personal communications with Yet-Ming Chiang of MIT suggest that a current mid-sized sedan, weighing about 3,000 pounds, requires roughly 300 watt-hours of battery capacity for every mile of range. This figure for a mid-sized sedan is roughly comparable to the 2011 Nissan Leaf, which weighs 3,354 pounds. The Leaf has a 24 kilowatt-hour battery pack and has an range rating of 73 miles from the Environmental Protection Agency, which translating to 328 watt-hours per mile. A mid-sized sport-utility vehicle, weighing roughly 4,000 pounds, requires 425 watt-hours for every mile of range. For a 200-mile range, which is significantly lower than current internal-combustion-based vehicles, the mid-sized sedan would require a 60 kilowatt-hour battery pack, while the mid-sized sport-utility vehicle would require 85 kilowatt-hour battery pack. As a third point of reference, a 2011 Ford F-150 SuperCab weighs 5,500 pounds. If the relationship is roughly linear, a pickup truck of this size would require a 123 kilowatt-hour battery pack. I should note that I am ignoring the effects of the battery's weight, which have real consequences (Kromer and Heywood (2007)). For example, the battery and control module for the Nissan Leaf weighs over 600 pounds.

A report from the [National Research Council \(2010\)](#) estimated current battery costs and projected future costs for plug-in hybrid vehicles. The committee set the most probable current cost for a battery at \$875 per kilowatt-hour, with \$625 per kilowatt-hour being an optimistic estimate. They project battery costs falling by 35 percent by 2020 and 45 percent by 2030. At these prices and assuming they scale up to the larger battery sizes required for all-electric vehicles, currently

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<sup>5</sup>The efficiency of current electric motors is roughly 80 percent efficiency—meaning 80 percent of the energy in electricity goes to moving the vehicle, while current internal combustion engines are in the low 20 percent range. The theoretical bound on efficiency is roughly 30 percent for the internal combustion engine. For a reasonably accessible explanation, see [Johnson \(2003\)](#).

the battery alone for a mid-sized sedan with a range of 200 miles would cost between \$38,000 to \$50,000; the cost of a battery for the mid-sized sport-utility vehicle would be \$53,000-\$70,000, and a battery for the F-150 would cost between \$76,000-\$101,000. The optimistic values in 2030 for battery costs alone would be \$21,000 for the sedan, \$29,000 for the sport-utility vehicle, and \$42,000 for the full-sized pick-up truck. The lower cost per mile of electric vehicles would offset these higher upfront costs to some extent. The sedan, for example, at average retail electricity rates would cost 3 cents per mile, compared to roughly 13 cents per mile at a gasoline price of \$4/gallon and a fuel economy of 30 miles-per-gallon. However, these savings in operating costs are unlikely to outweigh the upfront costs at any reasonable discount rate (Anderson (2009)).

While all-electric vehicles may not be cost competitive, vehicles that are partly propelled by electricity, such as hybrids or plug-in hybrids, may be. Hybrid and plug-in hybrid vehicles is that they economize on these battery costs because they use a higher share of the battery's capacity for typical driving patterns. Put another way, if a consumer could size the battery in an all-electric vehicle for each specific trip, all electric vehicles may be cost competitive at current battery prices. To underline this point, Anderson (2009) calculates that a plug-in hybrid with a 10-mile range is cost competitive even at battery costs of nearly \$2,000 per kilowatt-hour. Similar themes are echoed in the more comprehensive analysis of Michalek et al. (2011).

The National Research Council (2010) battery cost estimates are somewhat controversial. The estimates accord well with the published cost estimates for the Nissan Leaf's battery of \$750 per kilowatt-hour (Loveday, 2010) and are within the range of estimates I have seen for the Chevrolet Volt's 16 kilowatt-hour battery pack (\$500-\$930) per kilowatt-hour (Wright, 2011; Peterson, 2011). However, a number of industry trade groups argue that their costs are too high (for example, Electrification Coalition (2009), CalCars (2010)). Better Place, a swappable electric vehicle battery company, has stated that they are purchasing batteries at \$400 per kilowatt-hour. Other studies estimate much lower prices under hypothetical situations. For example, Nelson, Santini, and Barnes (2009) and Amjad, Neelakrishnan, and Rudramoorthy (2010) simulate battery costs as low as \$260 per kilowatt-hour using engineering models of production. These results rely heavily on large scale economies and an assumption that plants operate 24 hours a day. Under these assumptions, costs fall by as much as an order of magnitude when production increases from 10,000 to 100,000 units per year. Figure 4 plots a number of battery cost estimates over time from a recent review article as well as the goal of the United States Advanced Battery Consortium (Cheah and Heywood (2010)): clearly, the estimates show a large dispersion in all years.

The true cost of batteries, both now and certainly in the future, is unresolved. But these calculations suggest that some major technological breakthrough may be needed for electric vehicles

to play a large role in reducing oil consumption: either a much lower-cost battery, or technological breakthroughs that allow reductions in the size and/or weight of vehicles, perhaps through the use of polymer, aluminum, or composite body panels. However, technological breakthroughs reducing size and weight could also be applied to internal combustion engines, and could thus have significant effects on oil use in that way—without leading to greater use of electric cars (Knittel (2011)). Alternatively, the ranges of electric vehicles could end up being much shorter than we are accustomed to hearing about. Indeed, the battery-powered Nissan Leaf is rated at a range of 73 miles. Air conditioning or heating—because heat from the internal-combustion engine can no longer be used to heat the interior of the car—significantly reduces this range. Car and Drivers road test for the Nissan Leaf finds an average range of 58 miles and discusses the effect of heating (Gluckman (2011)).

Hydrogen vehicles also take advantage of the higher efficiency inherent in electric motors, but generate their own electricity via a fuel cell. Support for hydrogen vehicles has significantly waned over the past decade, but pursuing the possibility of a hydrogen-fueled car remains a stated objective of the U.S. Department of Energy. Hydrogen vehicles use a fuel cell, which uses a proton exchange membrane to convert stored hydrogen and oxygen from the surrounding air into electricity; the by-product of this conversion is water. Fuel cells are cheaper than batteries and refueling could be much faster. (Although supporters of batteries sometimes argue that you could refuel quickly via a system of swappable batteries.) At present, however, hydrogen refueling is not simple, with some stations requiring special suits and apparatus.

Detractors of hydrogen vehicles often point to the fact that they are far less efficient than electric vehicles on a well-to-wheel basis; that is, they take more total energy to travel one mile, because of the energy needed in making the hydrogen. However, the more relevant question is relative costs of the two technologies. That is, if the added energy needed to produce hydrogen were free or low-cost to society, then the added inefficiency would not matter or would matter less. That is not to say hydrogen vehicles do, in fact, have lower costs. For hydrogen vehicles, the relevant costs are: the cost of the fuel cell, the cost of the high-pressure storage tank, the cost of hydrogen, and infrastructure costs.

The first main cost element for a hydrogen-fueled cars are the fuel cells, which are currently quite expensive. A recent U.S. Department of Energy study (James, Kalinoski, and Baum (2011)) estimates that the cost of fuel cells at the current fairly low production levels are roughly \$230 per kilowatt. To understand what this means for costs, the Chevy Volt has a 111 kilowatt-electric motor, while the Nissan Leaf has a 80-kilowatt motor. The Volts motor is equivalent to a 149 horsepower engine, which is about the amount of horsepower from a four-cylinder gasoline engine.

Manufacturers appear to install fuel cells equivalent to the size of the motor, so the Volt would require a 111-kilowatt fuel cell at a cost over \$25,000. (There is a prototype Toyota Highlander FCV on loan to the University of California-Davis that combines a same-sized motor and fuel cell. The Honda FCX Clarity does so as well.) The alternative is to hybridize the vehicle by combining a fuel cell with a rechargeable battery back-up. Of course, electric motor and fuel cell combinations with horsepower levels comparable to larger vehicles would need to be correspondingly much larger.

As with some of the literature on battery costs, a number of papers on the future costs of fuel cells are built on assumptions of large-scale economies. Using engineering-economic simulations models, the U.S. Department of Energy study uses a scale economy elasticity of -0.2, and thus simulates that a fuel cell manufacturer producing 500,000 units per year could do so at an encouraging cost of \$51 per kilowatt (James, Kalinoski, and Baum (2011)). Given the importance of this possibility of gains from economies of scale and learning-by-doing, more studies along these lines seems like an important area for future research.

The second major cost component for a hydrogen vehicle is the storage tank. Hydrogen is ideally stored as a liquid under pressure, because this has the highest energy density. BMW recently demonstrated a hydrogen vehicle with liquid storage. However, storing hydrogen as a liquid faces major obstacles, as the National Research Council (2004) study points out. For example, the liquid must be kept at -252 degrees Celsius, and the liquid storage tanks currently cost roughly \$500 per kilowatt-hour of energy stored, with the next generation perhaps dropping then cost to roughly \$100 per kilowatt-hour. (Brunner (2006)). Again using 60 kilowatt-hours as a reasonable guideline for a mid-sized sedan that can travel 200 miles, the storage tank alone would cost \$30,000 using current technology and \$6,000 using the projected next generations technology.

Thus, absent a major technological breakthrough in liquid storage, hydrogen is likely to be stored as a compressed gas, which either increases the space required for the storage tank or reducing the range of the vehicle (Ogden et al. (2011)). Costs of compressed storage tanks, if produced at a large scale, might fall between \$15 and \$23 per kilowatt-hour of energy (Ogden et al. (2011)). Therefore, the storage tank for 3000-pound sedan with a range of 200 miles would cost between \$900 and \$1,400. However, gas storage tanks face durability issues, which are addressed by making the tanks larger. Indeed, the volume of a tank of this size is large enough such that manufacturers are likely to design the vehicle around the tank (National Research Council (2004)). However, if the estimated scale economies truly exist for both fuel cells and storage tanks, the combined cost of the fuel cell and storage tank for a hydrogen vehicle have the potential to be much cheaper than the battery required for an electric vehicle.



The third component is the cost of the hydrogen fuel itself, often quoted in terms of dollars per kilogram. A kilogram of hydrogen has roughly the same energy content as a gallon of gasoline, and given the increased efficiency of the electric drive-train, it can propel the vehicle roughly twice as far as a gallon of gasoline (for discussion, see [National Research Council \(2004\)](#), Appendix H). Here, too, the engineering literature suggests the possibility of large scale economies. Hydrogen can be produced in many ways, ranging from on-site production facilities to larger facilities where hydrogen is then shipped to refueling stations. [Weinert and Lipman \(2006\)](#) provide engineering cost estimates of the long-run average cost of hydrogen. Cost estimates vary considerably, but are as low as \$4.90/kg. Accounting for the more efficient motors (and taxes on gasoline), this is roughly on par with current gasoline prices.

The current federal subsidy for electric vehicles is a tax credit of \$2,500 plus \$417 for each kilowatt-hour of battery capacity in excess of 4 kilowatt-hours, with a maximum tax credit of \$7,500. Both the Nissan Leaf and Chevrolet Volt qualify for the maximum tax credit. The Toyota Prius Plug-in Hybrid, with a battery size of 4.4 kilowatt-hours, qualifies for a \$2,500 tax credit. There is also a federal tax credit for installation of charging equipment equal to 30 percent of the cost, with a maximum tax credit of \$1,000 for residences and \$50,000 for businesses ([Belson \(2011\)](#)). A variety of state-level policies also exist with tax credits as high as \$6,000 for qualifying vehicles (in Colorado).

One open question is whether, given the apparent need for technological breakthroughs for either electric or hydrogen vehicles, the funds used for these subsidies would be better served subsidizing research and development. The battery industry points to a number of potential game changers, such as lithium-air batteries and semi-solid flow cell batteries. Lithium-air batteries have a much higher energy density compared to the lithium-ion batteries presently used in the Leaf and Volt, leading to as much as five to ten times more energy for a given weight than lithium-ion batteries and twice the energy for a given size ([Zyga \(2011\)](#)). However major hurdles exist. These batteries are prone to get clogged as lithium-oxide builds up in the battery, and therefore cannot be recharged as often as would be needed in a vehicle. Semi-solid cell batteries suspend the positive and negative electrodes in a liquid electrolyte ([Chandler \(2011\)](#)). This not only has the potential for efficiency gains, but the battery can also, in principle, be refueled by draining the spent liquid and pumping in full-charged liquid. This battery structure is still in its infancy, however.

## 5 The Forgotten Channel: Reductions in Vehicle-Miles Travelled

The final channel for reductions in oil consumption is reductions in vehicle-miles travelled. U.S. energy policy has largely ignored this channel. Indeed, policies like the Corporate Average Fuel Economy standards and biofuel subsidies push in the opposite direction, in the sense that they reduce the marginal cost of driving an extra mile.

Figure 5 plots per capita vehicle miles-travelled in the U.S. from 1970 to 2009. The general trend upward is remarkable, with vehicle-miles travelled nearly doubling from 1970 to 2008. Remember that the figure shows per capita growth in vehicle-miles travelled, so that total growth in vehicle-miles travelled, including that attributable to population growth, would be even more striking. The figure also graphs real oil prices on the right-hand axis. The two price spikes in real oil prices—in the late 1970s and early 1980s, and in the last few years—are clearly correlated with a flattening out of vehicle miles travelled. Conversely, the period of dropping oil prices over much of the intervening period is a time when vehicle-miles travelled soared. Of course, this connection is only illustrative: a full analysis of how the price of oil affects vehicle-miles travelled would need to make additional adjustments for changes in income, business cycles, and more. But more detailed analyses do offer strong evidence that vehicle-miles travelled does respond to gasoline prices.

The response of vehicle-miles travelled to changes in gasoline prices varies, as one might expect, by the time frame for adjustment. The short run offers little scope for reductions in vehicle-miles travelled, and so the measured elasticity is likely to be small. For example, [Small and Dender \(2007\)](#) estimate that one-month elasticity of vehicle-miles travelled to changes in price was -0.02 between 1997 and 2001, with a similarly calculated short-run elasticity of -0.05 from 1966 to 2001. [Hughes, Knittel, and Sperling \(2008\)](#) estimate the one-month elasticity for use of gasoline—which is largely driven by the one-month elasticity of vehicle-miles travelled. They find that the one-month elasticity in the 1970s was roughly -0.3, while it has fallen to roughly -0.07 in the 2000s.

[Knittel and Sandler \(2011\)](#) estimate an elasticity of vehicle-mile travelled over two years using observations on vehicle odometers in California's smog check program. They estimate an average elasticity of between -0.16 and -0.25 (see also [Gillingham \(2011\)](#), who finds similar estimates). More importantly, they find that the dirtiest quartile of vehicles in terms of their criteria pollutants are over four times more responsive to changes in gasoline prices than the cleanest quartile, while dirtier vehicles in terms of their greenhouse gas emissions are roughly twice as sensitive. This increases the emission reductions resulting from a higher fuel prices.

Long-run estimates of the elasticity of vehicle-miles travelled with respect to price are more difficult to identify, given that no sustained price increase exists in the data. The literature has thus

focused on estimating partial adjustment models. This approach is necessarily imperfect, because if consumers believe the price change to be temporary, then the partial adjustment parameter leads to an underestimate of the long-run elasticity. Using a partial adjustment model, [Small and Dender \(2007\)](#) estimate a long-run elasticity of vehicle-miles travelled with respect to price of -0.11 from 1997 to 2001 and -0.22 across their entire sample from 1966 to 2001. A number of earlier studies exist that find roughly similar results to the short-run, medium-run, and long-run findings described here. [Graham and Glaister \(2002\)](#) and [Dahl \(1995\)](#) provide surveys.

Few existing policies seek reductions in vehicle-miles travelled, other than subsidies for public transit. [Parry and Small \(2009\)](#) present evidence that large public transit subsidies are welfare improving. The main benefit, however, arises through relieving congestion, not through a significant reduction in petroleum usage. Given the reluctance of policy makers to adopt Pigouvian taxes that would affect petroleum consumption, this approach to reducing petroleum use is likely to be underutilized.

## 6 Discussion: Pigouvian Taxes and Policy Choices

When economists are confronted with negative externalities, their trained reaction is that economic actors need an incentive to take the social costs of their actions into account in their decision at the margin. This outcome can happen through a Pigouvian tax, a system of tradeable permits, or a system of clarified property rights. Here, I use the Pigouvian tax—in this case, a tax on petroleum or greenhouse gas emissions that would include the value of the environmental and other externalities discussed at the start of this article—as a benchmark with which to compare other policy options for reducing U.S. petroleum use and reduce greenhouse gas emissions.

Absent other market failures, it is clear that performance standards, such as Corporate Average Fuel Economy standards and Renewable Fuel Standards, will be less efficient than Pigouvian taxes in curbing gasoline consumption. Most basically, performance standards act as an implicit tax and subsidy program. Any product better than the standard is implicitly subsidized, while any product worse than the standard is implicitly taxed. In the case of fuel-related policies, such as the Renewable Fuel Standard, this implies an implicit subsidy for fuels that are greener but nonetheless emit greenhouse gases, driving a wedge between the Renewable Fuel Subsidy and the efficient policy.

Additional inefficiencies exist with respect to Corporate Average Fuel Economy standard. At a basic level, it focuses on the wrong thing—fuel economy instead of total fuel consumption. CAFE only targets new vehicles and leads to subsidies for some vehicles. Finally, CAFE pushes consumers

into more fuel-efficient vehicles without changing the price of fuel, leading to more miles travelled. The empirical size of this last effect, known as rebound, is a matter of ongoing research, but to the extent that rebound occurs, it necessarily leads to greater congestion, accidents, and criteria pollutant emissions relative to the status quo. These added externalities loom even larger, because the first-best outcome would lead to reductions in vehicle-miles travelled, not increases.

The Corporate Average Fuel Efficiency standards have often been analyzed in comparison with a Pigouvian tax. For example, [Kleit \(2002\)](#) investigates the long-run effects of a 3 miles-per-gallon increase in the CAFE standard and the gasoline tax that would achieve the same reduction in consumption of gasoline. He finds that the CAFE standard leads to a \$3 billion per year social cost. An 11-cent gasoline tax achieves the same 5.1 billion gallon reduction in annual gasoline use at a social cost of \$275 million. [Austin and Dinan \(2005\)](#) find similar results. They simulate the costs of a 3.8 miles-per-gallon increase in CAFE and the required gas tax that achieves the same gasoline reductions over the 14 years in which the change in CAFE becomes fully implemented. They find that CAFE is between 2.4 and 3.4 times more expensive than the equivalent gasoline tax. The average cost of an increase in CAFE of 3.8 miles-per-gallon, on a per ton of carbon dioxide abated basis, is between \$33 and \$40. At the social cost of carbon estimates of [Greenstone, Kopits, and Wolverton \(2011\)](#), discussed earlier, this suggests that increases in CAFE of this magnitude either reduce or have a small increase in welfare.

[Jacobsen \(2011\)](#) estimates the relative efficiency of Corporate Average Fuel Economy standards and gas taxes and also focuses on the differential impacts of CAFE across U.S., Asian, and European automakers. He finds that CAFE is over seven times more expensive than a gasoline tax that achieves the same reductions in greenhouse gas emissions over the first 10 years. When Jacobsen allows for the technology to change in offered vehicles, he finds that the cost of CAFE falls by over 60 percent, but are still over twice that of the costs of the gasoline tax without technology options. Even after allowing for technology adoption, Jacobsen estimates that the cost of a one-mile-per-gallon increase in CAFE is over \$220 per ton of carbon dioxide saved, well above current estimates of the social cost of carbon.

A parallel line of research has compared subsidies and mandates for biofuels with a Pigouvian tax approach, again finding that the Pigouvian tax is much more cost-effective. For example, [Holland et al. \(2011\)](#) simulate the relative efficiency of ethanol subsidies and the Renewable Fuel Standard compared to a Pigouvian carbon tax using feedstock-specific ethanol supply curves meant to represent cost conditions in 2020. They first simulate the greenhouse gas reductions from the current subsidies and Renewable Fuel Standard in 2020, and find reductions of 6.9 and 10.2 percent from subsidies and the Renewable Fuel Standard, respectively. They then calculate the required

carbon tax that achieves a 10.2 percent reduction in emissions. Their results suggest that the social cost under subsidies is four times greater than the average social cost under the carbon tax, \$82 per ton of carbon dioxide under the subsidies compared to \$19 per ton for the carbon tax, despite the larger emission reductions under the tax. Similarly, the social cost under the Renewable Fuel Standard mandate is three times larger than the carbon tax, at \$49 per ton of carbon dioxide.<sup>6</sup> There are also unintended consequences associated with the fuel-based policies. Holland et al. shows that land use patterns vary considerably across subsidy and mandate programs, relative to the Pigouvian tax. If these land use changes exacerbate other negative externalities, such as fertilizer run-off or habitat loss, the inefficiency of these fuel-based programs will be understated.

The existing work on both Corporate Average Fuel Economy standards and the Renewable Fuel Standard suggest that increasing these policies may actually reduce aggregate welfare, given current estimates of the social cost of greenhouse gases. However, the research described thus far leaves out a potential market failure that can make fuel efficiency standards and biofuels subsidies appear at least somewhat more attractive: The question of whether consumers are myopic in their preferences about fuel economy. Myopic consumers may be unwilling to invest a dollar at the time of purchase for a savings of present-discounted value of a dollar some time in the future. Consumer myopia implies that correcting prices with a Pigouvian tax is not sufficient to achieve the first-best outcome. This opens the door for additional policies to complement Pigouvian taxes. If consumers apply a discount rate that is larger than their true discount rate when purchasing vehicles, then policies that alter their choices or the relative prices of vehicles can, in principle, raise welfare.

Several recent papers have explored this topic, and the paper by Allcott and Greenstone in this symposium takes up what they call the energy efficiency paradox in more detail. The evidence appears mixed. As one example, [Allcott and Wozny \(2011\)](#) find that, when purchasing a used vehicle, the average consumer discounts the future at a rate of 16 percent. In contrast, [Busse, Knittel, and Zettelmeyer \(2011\)](#) find no evidence of the energy paradox in the new car market and implied discount rates are most often below 12 percent in the used car market. There is also a possibility that consumers may be acting optimally in the sense that a 16 percent discount rate is on par with their cost of capital, but the optimal discount rate from a policymakers perspective may be much lower if, for example, society puts more weight on the welfare of future generations

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<sup>6</sup>The authors also find, however, that the alternatives to the carbon tax have the potential to yield large winners in low-populated counties—as high as \$6,800 per capita per year. They then find that Congressional voting on the Waxman-Markey cap and trade bill correlates with the simulated district gains and losses in predicted ways—Congressional members whose districts gain the most under the Renewable Fuel Standard were less likely to vote for Waxman Markey conditioning on the Congressional members political ideology, the districts per capita greenhouse gas emissions, whether the state is a coal mining state, and a variety of other potential determinants of voting behavior.

than a given consumer may choose to do.

If consumers are indeed myopic in their willingness to consider fuel efficiency, then second-best policies like CAFE standards counteract consumer myopia, as they push demand to more fuel-efficient vehicles. They do not, however, eliminate the need for Pigouvian taxes. Given the potential importance of consumer discounting for optimal policy in both transportation and in other sectors, this open question warrants further research.

How large must myopia be for Corporate Average Fuel Economy standards to be welfare-improving, in the absence of Pigouvian taxes? Fischer, Harrington, and Parry (2007) analyze the social cost of increases in Corporate Average Fuel Economy standards under two assumptions regarding consumer myopia: a) consumers mistakenly inflate their discount rate by 14 percent over the true discount rate of 4.5 percent; and b) consumers care only about the fuel costs in the first three year of the vehicles life, but their true discount rate is 4.5 percent. They consider the welfare implications of increasing CAFE in the presence of a variety of externalities, including local and global pollution, congestion costs, accident costs, and external costs associated with oil dependence. Their results suggest that CAFE standards are welfare-improving only when consumers account for the fuel costs in the first three years of a vehicles life. Increases in CAFE reduce aggregate welfare even when consumers undervalue the future by 14 percentage points. Welfare falls by roughly 20 cents per gallon of gasoline saved under this scenario. They do not consider the welfare costs of Pigouvian taxes under these different scenarios.

The results with respect to both Corporate Average Fuel Economy standards and the Renewable Fuel Standard underscore an important point: Second-best (or third-best) policies need not be welfare-improving in the presence of negative externalities. Continued work that helps us better understand in what circumstances they do improve welfare, and the magnitude of other market failures, is needed.

## 7 Conclusion

A policy that puts a price on the externalities, like a carbon tax or cap-and-trade policy, would be desirable in addressing the externalities created by petroleum fuels in the U.S. economy. But both because such policies seem impractical for political reasons and because of the possibility of consumer myopia, there is potentially a role to be played by supplementary policies. Given the current state of technology, biofuels, electric vehicles, and hydrogen fuel-cell vehicles remain some years in the future. Their eventual commercial viability probably depends on a combination of

technical breakthroughs, the emergence of economies of scale in production, continued high prices for gasoline, and policy.

Corporate Average Fuel Economy standards can play a useful role as a second-best policy, in pushing the developments in auto technology to pay more attention to fuel efficiency over horsepower and weight-adding ingredients. But just as the U.S. political system doesn't much like fuel taxes, it's worth noting that the political process found a way for the CAFE standards not to bind very much over the last few decades—by giving sports-utility vehicles and light trucks a lower standard, or even no standard at all. Furthermore, the literature calls into question whether increases in CAFE standards are welfare improving.

It will be interesting to see how the political system reacts if the significantly higher fuel economy standards planned for the next few years begin to bite for leading U.S. car manufacturers. It is worth considering some alternative second-best policies: for example, an open fuel standard that would require vehicles to be able to run on gasoline, ethanol, or methanol; gas guzzler-gas sipper feebate programs which mimic CAFE standards; or a vehicle-miles travelled tax. But ultimately, the single biggest influence on whether Americans reduce their consumption of petroleum-based fuels will probably be whether the forces of supply and demand in global markets that have kept oil prices relatively high since about 2005 continue to do so.

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## Tables

**Table 1:** Motor Fuel Taxes for OECD Category I Countries in 2010. (\$/gallon)

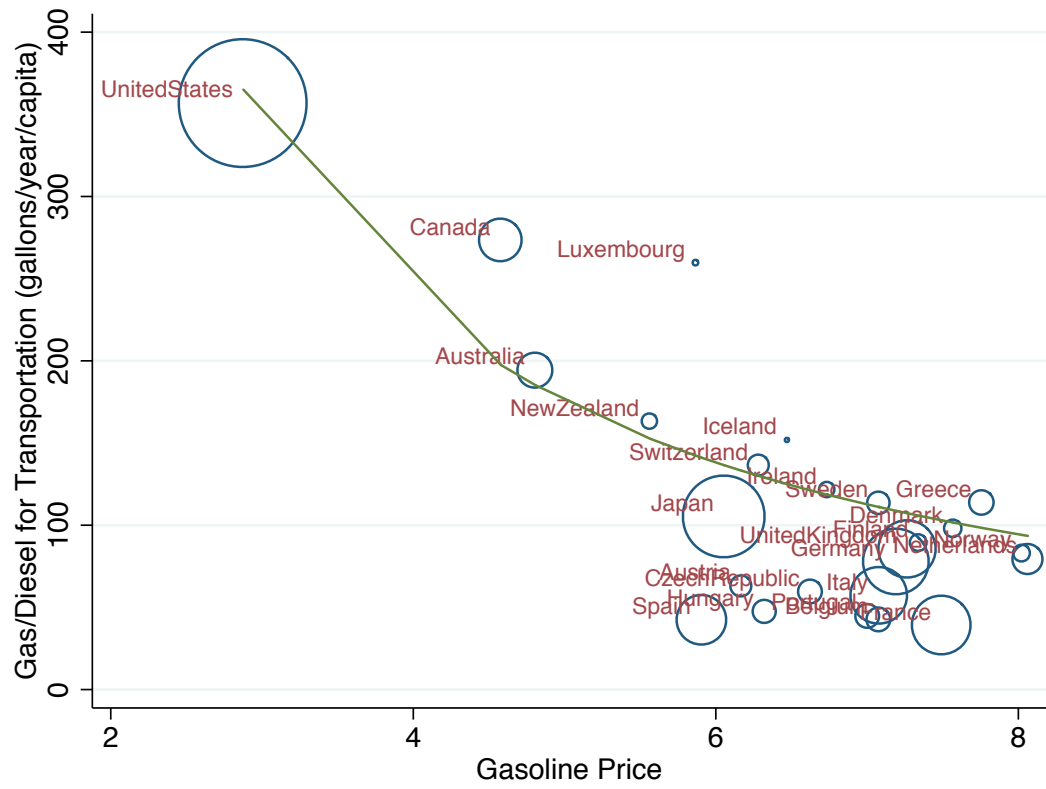
<b>Country</b>	<b>Gasoline</b>	<b>Diesel</b>
US	\$0.49	\$0.59
Canada	\$0.96	\$0.77
New Zealand	\$1.20	\$0.00
Australia	\$1.34	\$1.34
Iceland	\$2.28	\$2.03
Japan	\$2.59	\$1.55
Korea	\$2.64	\$1.87
Spain	\$2.66	\$2.08
Hungary	\$2.68	\$2.17
Austria	\$2.77	\$2.18
Luxembrg	\$2.90	\$1.94
Czech Republic	\$3.04	\$2.59
Switzerland	\$3.09	\$3.15
Slovak Republic	\$3.23	\$2.31
Sweden	\$3.24	\$2.56
Ireland	\$3.41	\$2.82
Italy	\$3.54	\$2.65
Belgium	\$3.58	\$2.10
Denmark	\$3.58	\$2.68
Portugal	\$3.65	\$2.28
France	\$3.80	\$2.69
Greece	\$3.82	\$2.39
Norway	\$3.87	\$2.97
Finland	\$3.93	\$2.28
UK	\$3.95	\$3.95
Germany	\$4.10	\$2.95
Netherlands	\$4.19	\$2.29

Data Source: Take from [http://www.afdc.energy.gov/afdc/data/docs/fuel\\_taxes\\_by\\_country.xls](http://www.afdc.energy.gov/afdc/data/docs/fuel_taxes_by_country.xls). AFDCs source is the OECD/EEA database on instruments for environmental policy, <http://www2.oecd.org/ecoinst/queries/index.htm>.

Notes: Rates as of 01/01/2010. Data for the United States and Canada include average excise taxes at the state/provincial level. VAT is not included.

# Figures

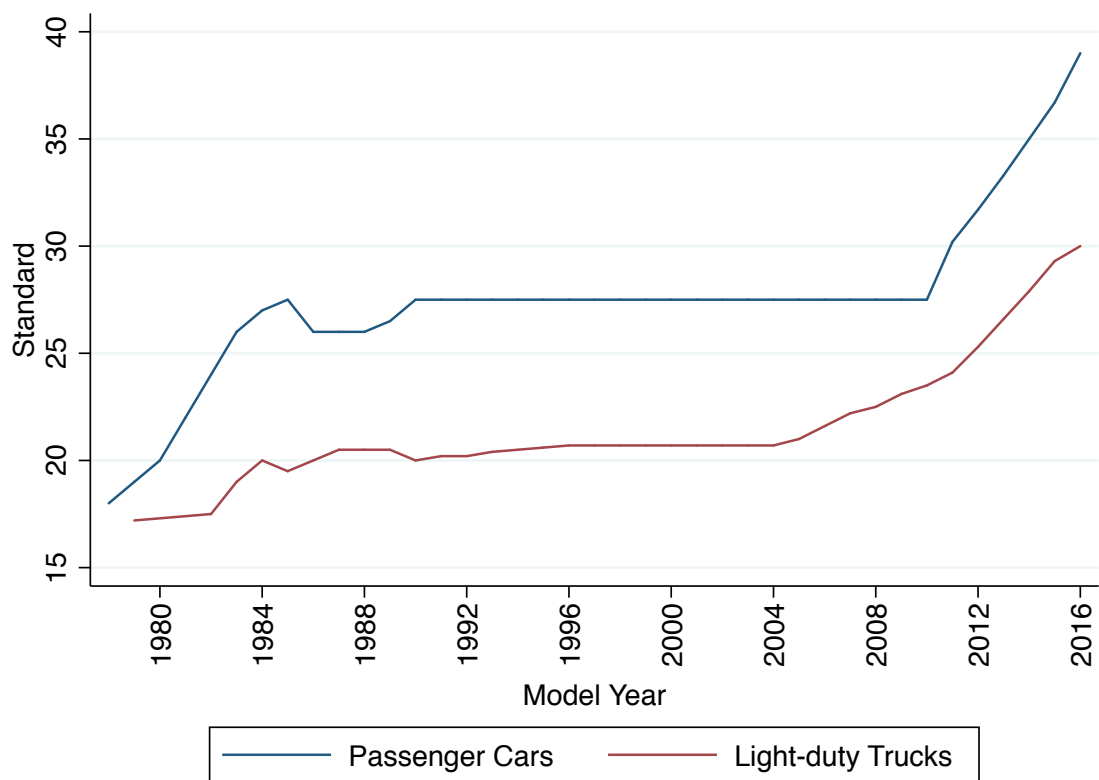
Figure 1: Transportation Fuel Consumer per Capita versus Fuel Price.



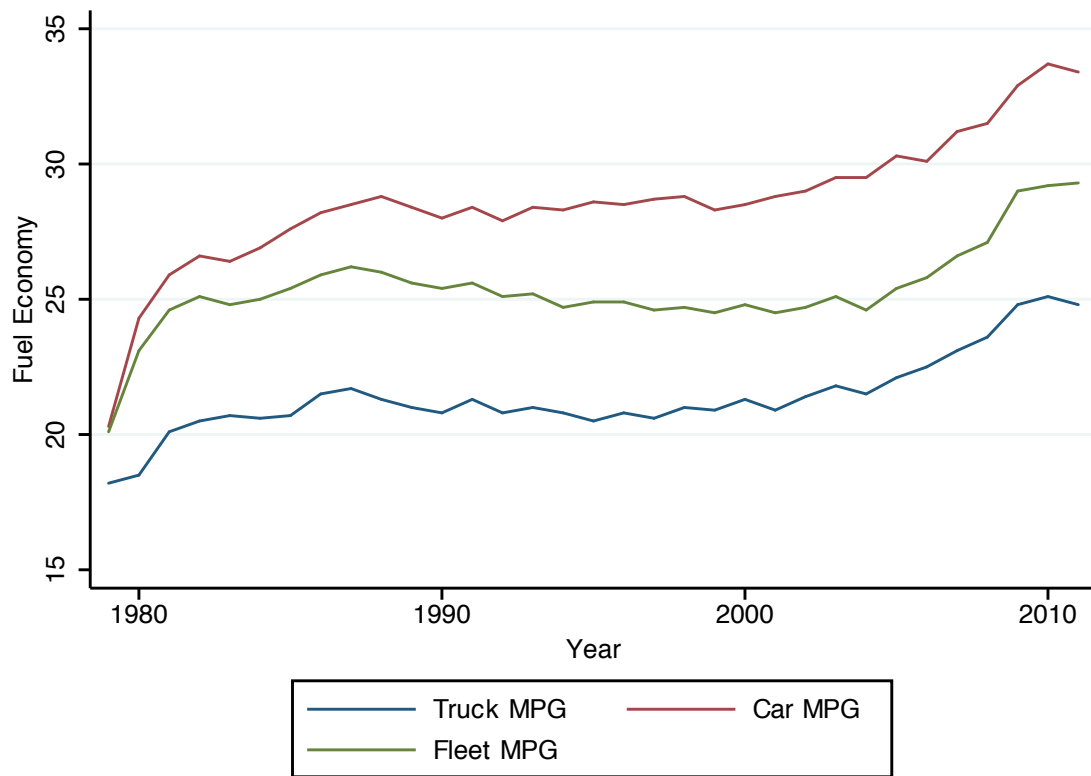
Data source: Worldbank.org

Notes: Size of the circle proportional to population. The line is the fitted value from a regression of the log of consumption on the log of price.

**Figure 2:** U.S. CAFE Standards from 1978 to 2016.

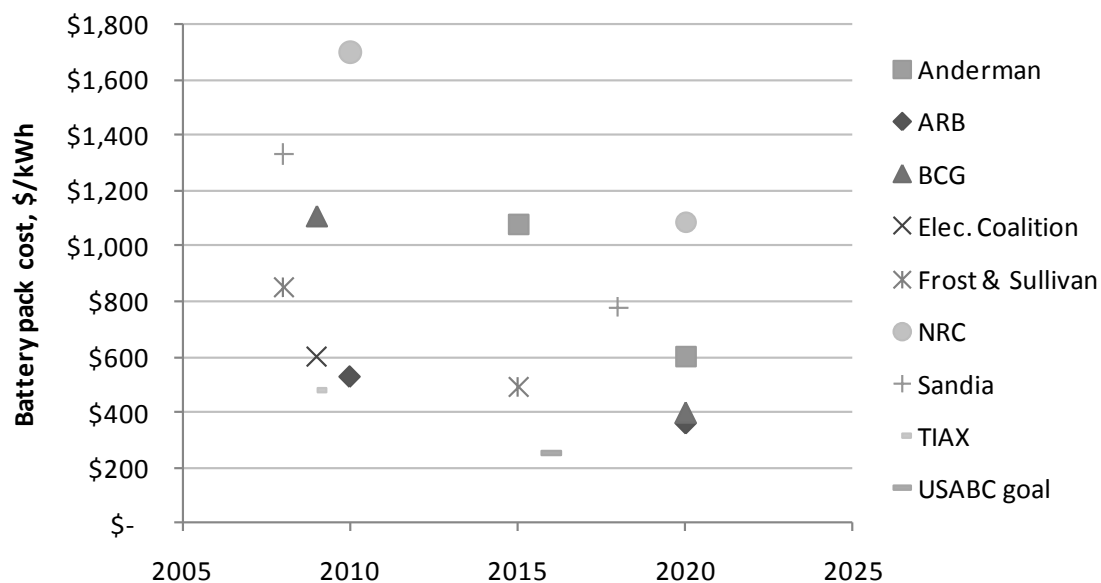


**Figure 3:** U.S. New Vehicle Fuel Economy from 1979 to 2011.



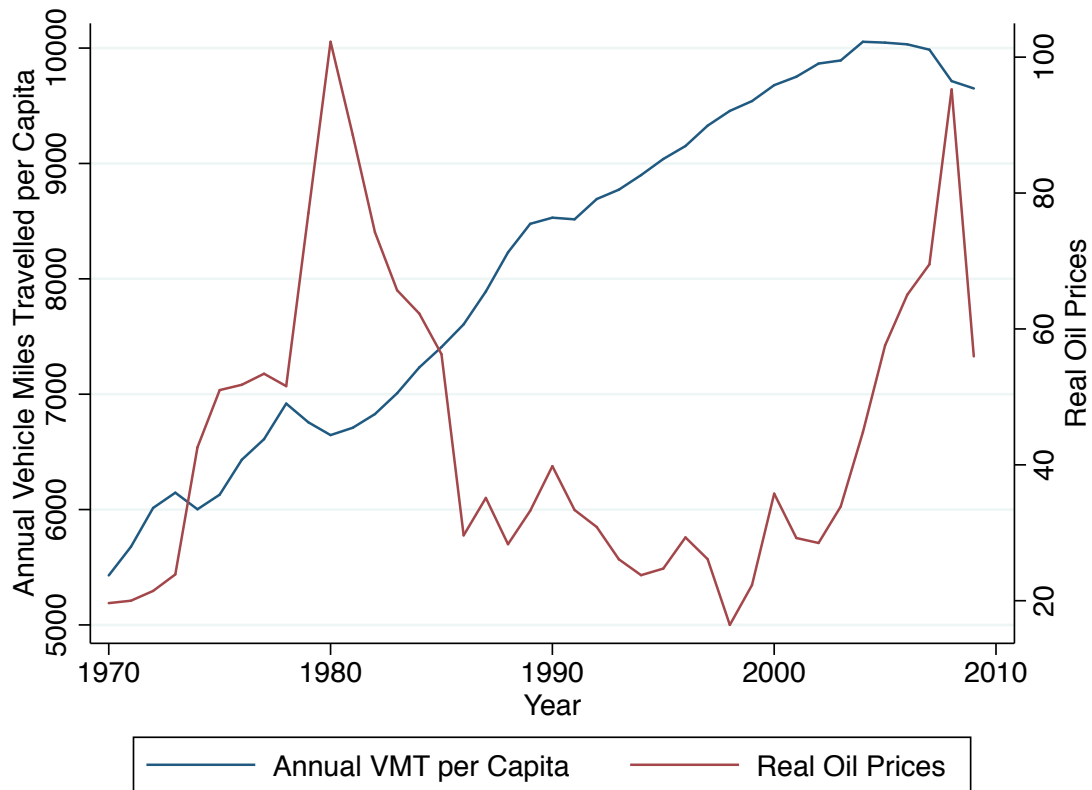


**Figure 4:** Battery Cost Estimates from the Literature (Cheah and Heywood, 2010).



Notes: Cost estimates from Anderman (2010), ARB (2009), Boston Consulting Group (2010) (BCG), Electrification Coalition (2009), Frost & Sullivan (2009), National Research Council (2010), Tons et al. (2008) (Sandia), Barnett et al. (2009) (TIAX), and Pesaran et al. (2007) (USABC). When a range is given in the original source, Cheah and Heywood plot the average. The USABC number is a goal, not a cost estimate.

**Figure 5:** Vehicle Miles Travelled per Capita from 1970 to 2009.



Data Sources: Oil prices are taken from the Energy Information Administration. Vehicle miles travelled are from the Department of Transportation, <http://www.fhwa.dot.gov/policyinformation/travel/tvt/history/>